Technical Support for the Bay-wide Runoff Reduction Method



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Section 1: Introduction and Objectives

The objective of this memo is to provide the scientific basis for creating a workable engineering framework to rapidly design effective combinations of runoff reduction and stormwater treatment practices to promote ESD within states and localities within the Chesapeake Bay watershed. This memo draws extensively from recent work performed by CSN and the Center for Watershed Protection to develop a state-wide compliance system for the Virginia DCR (Hirschman et al, 2008). Indeed, the Appendices are taken directly from that document, and the hard work of Dave Hirschman, Kelly Collins and other CWP staff are gratefully acknowledged. The runoff reduction framework described herein can be adapted to other Bay states, such as Delaware, the District of Columbia, Maryland, Pennsylvania and West Virginia as they develop new stormwater regulations, policies and design manuals. The basic method is flexible enough that each state can modify it to suit their unique conditions and water resources protection objectives.

Section 2: Basic Concepts

This section outlines the basic concepts that provide the technical foundation for the runoff reduction method. This section defines the treatment volume, runoff reduction volume, runoff reduction practices, nutrient EMC removal, Level 1 and 2 STP design, and the four step compliance process.

2a The Treatment Volume

The treatment volume is a slight variation of the 90% capture rule that was originally established in the MDE (2000) stormwater manual. The 90th percentile rainfall event is defined as one-inch of rainfall in most parts of the Bay watershed (although it can range from 0.9 to 1.2 inches. The treatment volume is defined by multiplying the 90th percentile rainfall depth by three site cover runoff coefficients present at the site (forest, turf, and impervious cover), as shown in the equation below.

$$Tv = P * (Rv_I * \%I + Rv_T * \%T + Rv_F * \%F) * SA$$

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Where

- Tv = Runoff reduction volume in acre feet
- P = Target Rainfall Depth
- Rv_I = runoff coefficient for impervious cover
- Rv_T = runoff coefficient for turf cover or disturbed soils
- Rv_F = runoff coefficient for forest cover
- % I = percent of site in impervious cover
- %T = percent of site in turf cover
- %F = percent of site in forest cover
- SA = total site area, in acres

Table 1. Site Cover Runoff Coefficients				
Soil Condition	Runoff Coefficient			
Forest Cover	0.02 to 0.05*			
Disturbed Soils	0.15 to 0.25*			
Impervious Cover	0.95			
*Range dependent on or	iginal Hydrologic Soil Group (HSG)			
Forest A: 0.02	B: 0.03 C: 0.04 D: 0.05			
Disturbed Soils A: 0.15	B: 0.20 C: 0.22 D: 0.25			
Restored Soils A: 0.05	B: 0.06 C: 0.10 D: 0.12			

The site cover runoff coefficients to be used are provided in Table 1

The three runoff coefficients provided in Table 1 were derived from research by Pitt et al (2005), Gregory et al (2004), Lichter and Lindsey (1994), Schueler (2001a), Schueler, (2001b), Legg et al (1996), Pitt et al (1999), Schueler (1987) and Cappiella et al (2006). Numerous researchers have documented the impact of construction earthworks on the compaction of soils, as measured by an increase in bulk density, a decline in soil permeability, and an increase in the runoff coefficient. These areas of compacted pervious cover (lawn or turf) have a much greater hydrologic response to rainfall than forest or pasture. The effect of earthworks and soil compaction nearly doubles the runoff coefficient from un-forested areas (as shown in Table 1).

The proposed treatment volume has several distinct advantages when it comes to evaluating and sizing runoff reduction and stormwater practices, including:

- Storage is a direct function of impervious cover and disturbed soils, which provides designers incentives to minimize the area of both at a site.
- The treatment volume can be set for water quality and/or channel protection volume rainfall depth, depending on the characteristics of the receiving stream and the intensity of development in its subwatershed. This avoids the segregation of WQ and CPv requirements that is inherent in many state stormwater manuals, and should reduce compliance costs, at low intensity sites.
- Provides adequate storage to treat pollutants for all storm events, which is important since the first flush effect has been found to be modest for many pollutants (Pitt et al 2005).
- Allows for all structural and non-structural practices to be assessed on a common basis according to a roof to stream sequence.
- Explicitly acknowledges the difference between forest and turf cover and disturbed and undisturbed soils, which creates incentives to conserve forests and reduce mass grading and provides a defensible basis for computing runoff reduction volumes for these actions.

2b The Runoff Reduction Volume

The runoff reduction volume is the primary stormwater treatment strategy to maintain the same predevelopment runoff **volume** delivered to the stream after site development. In its simplest terms, this means achieving the same predevelopment runoff coefficient for each storm up to a defined rainfall event. Runoff reduction (RR_v) is defined as the total runoff volume reduced through canopy interception, soil infiltration, evaporation, rainfall harvesting, engineered infiltration, extended filtration or evapo-transpiration. Extended filtration includes bioretention or dry swales with under drains that delay the delivery of stormwater from small sites to the stream system by six hours or more. The RRv is considered to be fully inclusive of the T_v at a development site, and designers are strongly encouraged by combining runoff reduction and stormwater treatment practices in a series to maximize the degree of runoff and pollutant reduction achieved at a site.

The runoff reduction strategy has many benefits when it comes to managing stormwater at a site. For example, runoff reduction:

- Eliminates the use of credits as originally set forth in the Maryland stormwater manual (MDE, 2000), and instead, makes use of runoff reduction practices an integral element of on-site compliance.
- Provides an objective measure to measure the aggregate performance of environmental site design, runoff reduction practices and stormwater treatment practices (STPs) together using a common currency (i.e., the treatment volume).
- Mimics predevelopment hydrology with respect to runoff volume, duration and velocity which is important to reduce the increased frequency and duration of runoff events that stream channels experience after development. If the Channel Protection rainfall depth is used to define the treatment volume, it provides an attractive alternative to the 24 hour extended detention Channel Protection volume first advanced in the 2000 MDE manual. Even if ED is still needed to comply, the runoff reduction volume used for water quality can be directly subtracted from CPv.
- Helps maintain groundwater recharge that supports stream baseflow when it is not raining, without specifying a mandatory infiltration or recharge requirement that can be problematic in certain terrain or soil conditions in the Bay watershed
- Enhances the degree and reliability of pollutant mass removal for runoff reduction and stormwater treatment practices since pollutant loads are the product of both stormwater flow volume and the treated pollutant concentration leaving a practice.

2c. The List of Runoff Reduction Practices

The following practices are considered to have runoff reduction potential:

- 1. Sheetflow to Conserved Open Space (formerly a stormwater credit)
- 2. Rooftop Disconnection (formerly a credit)
- 3. Permeable Paving
- 4. Green Roofs
- 5. Grass Channel (formerly a credit)
- 6. Bioretention/Dry Swale
- 7. Wet Swale
- 8. Infiltration
- 9. Extended Detention (low)
- 10. Soil Amendments
- 11. Rain Tanks/Cisterns

The specific amount of runoff reduction achieved by these practices is outlined in Table 3 and Appendix A.

2d STP Nutrient EMC Reductions

The Core ESD Principles recommend that Bay state stormwater manuals should contain specific and numeric performance criteria to assure the aggregate nutrient load delivered to the Chesapeake Bay and Coastal Bays from urban development is actually reduced over time (CSN, 2008). The recommended nutrient based limit for post-development phosphorus is about 0.25 lb/acre/yr. The load limit creates an accountability mechanism to ensure development projects really meet watershed objectives to protect the Chesapeake Bay. The basic concept is that new development on non-urban land must not exceed the average load for non-urban land using effective stormwater practices in the watershed. The proposed nutrient requirement is similar to proposed nutrient stormwater regulations under consideration in the Commonwealth of Virginia. The Virginia load limits were computed using the Chesapeake Bay Model Tributary Strategy Confirmation Runs. The load limits were established as the total of forest, crops, pasture and mixed open space, adjusted for delivery to the Bay (3,418,105 lbs for TP), divided by their total land area in the state (12,209,171 acres). This yields an average load of 0.28 lbs/ac/yr for TP and 2.68 lbs/ac/yr for TN.

A simple spreadsheet calculation is used to define on-site compliance with the nutrient load, using both runoff reduction and phosphorus removal rates for each practice.

Considerable analysis was performed to define the appropriate measure of nutrient removal by runoff reduction and stormwater treatment practices. Recent work has shown that it is extremely important to segregate out the reduction in event mean concentration (EMCs) as it travels through the practice. To this end, the most recent version of the NPRD was reanalyzed to define the median and 75% quartile reduction in EMC for both phosphorus and nitrogen for a range of practices (see Appendix B). Given the similarity

in EMC reductions for both phosphorus and nitrogen within the same practice, a decision was made to rely on a single nutrient (phosphorus).

2e Level 1 and Level 2 STP Design

Most state stormwater manuals in the Bay contain one basic set of design criteria for the major groups of STPs, with some design features that required, with others that are merely encouraged. The runoff reduction method departs from this by establishing a baseline design level (Level 1) and a more sophisticated design (Level 2) that can achieve a greater degree of runoff and pollutant reduction. This is important since research has shown that the performance of ESD and STP practices can vary greatly depending on both sizing and design features.

Section 3 describes the technical approach that was undertaken to tease out which design and sizing factors could be assigned to baseline Level 1 design and more innovative designs that would have greater runoff and nutrient reduction (Level 2).

2f Summary Table for Current Practices

Table 2 summarizes the comparative runoff reduction, EMC reduction and total mass reductions associated with different ESD practices, based on research documented in Appendix A and B. When a range is shown, the first number is for Level I STP design and the second is for Level 2 STP design. In general, the practices fall into three groups (1) practices that are effective in both runoff and pollutant reduction (2) practices that are effective in runoff reduction but not pollutant reduction, and (3) practices that are ineffective in runoff reduction but are effective in pollutant reduction.

Table 2: Comparative Runoff Reduction, TP and TM EMC Removal Rates					
Practice	Level 1 RR	Level 2 RR	TP EMC PR	TN EMC PR	
	(%)	(%)	(%)	(%)	
Infiltration	50	90	25	15	
Bioretention	40	80	25 to 50	40 to 60	
Pervious Paver	45	75	25	25	
Green Roof	45	60	0	0	
Dry Swale	40	60	20 to 40	25 to 35	
Rain Tanks/Cisterns	40	40	0	0	
Rooftop Disconnection	25	50	30	15	
Grass Channel	15	30	15	20	
Dry ED Pond	0	15	15	10	
Wet Pond	0	0	50 to 75	30 to 40	
Constructed Wetland	0	0	50 to 75	25 to 55	
Sand Filter	0	0	60 to 65	30 to 45	
See Appendices A, B and C	for the derivation	n of these num	bers		

2f. The Four Step Compliance Process

The runoff reduction method relies on a four step compliance procedure, as shown in Figure 1, and described below.

<u>Step 1: Apply ESD Practices to Minimize IC, Grading and Loss of Forest Cover</u>. This step focuses on how to implement environmental site design practices prior to site layout using the ESD plan and map. The goal is to minimize impervious cover and mass grading, and maximize retention of forest cover, natural areas and undisturbed soils.

<u>Step 2: Compute Post Development Land Cover</u>. In the second step, a spreadsheet is used to compute the three site runoff coefficients, and calculate a site-specific target treatment volume and phosphorus load reduction limit.

<u>Step 3: Apply Runoff Reduction Practices</u>. In this step, the designer experiments with combinations of the nine runoff reduction practices on the site. In each case, they estimate the spatial area to be treated by each runoff reduction practice, and chip away at the required treatment volume within each drainage area on the site.

<u>Step 4: Compute Phosphorus Reduction by RRPs and Conventional STPs</u>. In this step, the spreadsheet checks to see whether the phosphorus load reduction has been achieved at the site. Removal by previously entered runoff reduction practices is automatically calculated, and the designer can then add conventional STPs such as filtering practices or linear wetlands to meet their remaining requirement, if needed.

In reality, the process is iterative for more difficult sites. When compliance cannot be achieved on the first try, designers can return to prior steps, to see what alternative combination of ESD practices, runoff reduction practices, or conventional STPs are needed for compliance. In the event that compliance is impossible on the site, the spreadsheet computes the unmet phosphorus load for the site, which would be subject to an impervious cover mitigation fee.

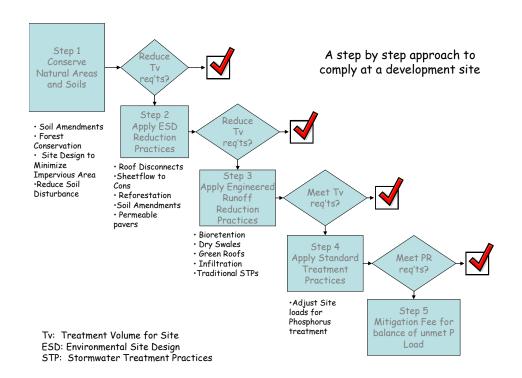


Figure 1. Sequence for Assessing Runoff Reduction Opportunities at a Site

Section 3. Basis for Level 1/Level 2 STP Design Guidelines

This section documents the scientific rationale and assumptions used to assign sizing and design features to Level 1 and Level 2 STPs that are presented in Appendix C.

Standard Design Features. The first step identified the "standard" design features that should be included in all designs (i.e., not directly related to differential nutrient removal rates). These include any features needed to maintain proper function of the STP, as well as its safety, appearance, safe conveyance, longevity, feasibility constraints, standard setbacks or maintenance needs. These standard features will be outlined in the detailed design specifications to be developed by CSN and others later in the year.

Design Point Tables. Appendix B of the Stormwater Retrofit Manual (Schueler et al, 2007) contains a series of tables that describe design factors that increase/decrease overall pollutant removal rates, and these were initially used to assign design features into Level 1 and 2. It should be acknowledged that design point method was primarily developed to evaluate removal rates for under-designed retrofits that may lack the full range of design features present in a new development setting. For example, the method evaluates removal for quartiles above and below the median removal. For a new development setting, the base removal rate is the median (i.e., Level 1), whereas Level 2 is the 75th percentile value. In addition, the original design point method was designed to estimate

removal for eight different pollutants; changes were made in this memo to reflect the more specific goal of nutrient removal.

Review of 2007 NPRD Rates, CWP (2007) recently released an update to the Winer (2000) national pollutant removal database. 27 new performance monitoring studies were added, mostly for under-represented practices such as bioretention, infiltration and water quality swales. Even so, nearly 80% of the performance entries in the NPRD were built and monitored from 1980 to 2000, so many of the older designs may not reflect modern design features (particularly for ponds and wetlands).

Review of Individual Studies. To gain additional insight into the value of different sizes and design features, 50 stormwater technical notes were reviewed that provided a more in-depth analysis of more than 70 studies included in the NPRD (Schueler and Holland, 2000). In addition, selected references were reviewed from the 2000 to 2006 stormwater literature, with an emphasis on design enhancements for infiltration, bioretention, and water quality swales. Greater emphasis was placed on studies in close geographic proximity to the Bay states.

Based on the foregoing analysis, five primary design factors were used to define Level 1 and Level 2 STPs: increased treatment volume, increased runoff reduction volumes, enhanced design geometry, vegetative condition, and use of multiple treatment methods. More on the basis for each split are provided below.

1. Increased Treatment Volume. Increasing the treatment volume can enhance nutrient removal rates, up to a point. The existing treatment option captures about 90% of the annual runoff volume, so further increases can only bring out modest increases, unless the larger volume increases the residence time or rate of nutrient uptake (which has been documented for ponds and wetlands). Therefore, three incremental levels of greater treatment volume were considered for each STP: 110%, 125 and 150% of the base Tv.

2. *Increased Runoff Reduction Volume*. The second strategy to enhance nutrient removal rates is to increase the proportion of the treatment volume that is achieved by runoff reduction. In this instance, design features that could significantly enhance runoff reduction volumes were generally assigned to Level 2.

3. Enhanced Design Geometry. A third strategy to split STPs according to nutrient removal is to isolate geometry factors that are known to influence either hydraulic performance or create better treatment conditions. Examples include flow path, depth of filter media, multiple cells, the SA/CDA ratio, max CDA and minimum ED time.

4. *Vegetative Condition*. A fourth splitting strategy involves the ultimate type and cover of vegetation within the STP insofar as it influences nutrient uptake, increases the ET pump, or stabilizes trapped sediments or a filter bed. Landscape

designs that maximize tree canopy or otherwise increase the ultimate vegetative cover for a practice were often used to support Level 2 designs.

5. *Multiple Treatment Methods*. The last major strategy is to combine several treatment options within a single practice to increase the reliability of treatment.

Section 4. Deriving Minimum Design Criteria for Select ESD Practices

Under the runoff reduction method, several ESD practices that were originally offered as credits in some state manuals (e.g., MDE, 2000) are now full fledged runoff reduction practices. From a design standpoint, however, it is still important to establish qualifying criteria for the following practices:

- Site Reforestation
- Soil Restoration
- Sheetflow to Conserved Open Space
- Rooftop Disconnection
- Grass Channels

The updated design criteria for these ESD practices are provided in Appendix D. In most cases, the new design criteria were based on the original qualifying credit criteria contained in the 2000 MDE Manual, but they have been updated to reflect local experience and further credit details in other manuals produced since 2000 (e.g., Minnesota, Credit River, DCR). The soil restoration and site reforestation criteria were drafted using recent research.

Section 5: The VA DCR Compliance Spreadsheet

A hard copy of the current version of the VA DCR compliance spreadsheet is provided in Appendix E. The spreadsheet is being tweaked based on input from design engineers and testing on difficult sites, and will be finalized in the summer of 2008. Several minor adaptations would need to be made if it were to be adopted for ESD compliance in other Bay states. Most engineers like the simplicity of the spreadsheet and particularly like the ability to quickly experiment with combinations of runoff reduction practices to find the most cost-effective solution. Plan reviewers like the fact that most of the calculation cells are hidden (and can't be changed by the user), so the focus is on coming up with the best combination of practices.

The spreadsheet method has been tested on about a half dozen highly constrained sites (existing site plans for small infill, flat coastal plain, highways, and steep slope sites). In most, but not all cases, designers found acceptable solutions using the spreadsheet for water quality (one-inch). The ability to meet the channel protection criteria was not tested, since this sizing issue has yet to be resolved in Virginia.

Section 6 Review and Verification in Site Plan Review and Construction

Runoff reduction needs to be explicitly addressed during three stages of local development review- feasibility during concept design, confirmation in final design, and verification during final construction inspection at the site.

<u>1. Early Concept Design</u>. Practices are initially considered during site layout by carefully considering existing drainage features, forest conservation, stream buffers, wetland, floodplain, recharge, habitat, steep slopes, zero-order stream protection, and other natural area protections that apply to the site. The early map/plan should include initial estimates of site forest, turf and impervious cover, and the initial spreadsheet calculation indicating how they have met the treatment volume and pollutant removal requirements. The local review authority then checks both the practice delineations and the computations as part of the stormwater concept plan review.

<u>2. Final Design</u>. The practices are reviewed a second time during final design to confirm whether they meet the both the spreadsheet and practice feasibility requirements (e.g., slopes, contributing drainage area, flow paths, etc). The designer should be able to justify the precise boundaries of each practice area drawn on the plan, and indicate in the submittal whether any additional grading, soil amendments or plantings are needed to qualify. Reviewers would check practice area delineations, make sure flow paths are realistic, and make sure any required easements or management plans needed for the practice are secured. It should be emphasized that full engineering review would still be required for any individual structural practice used at the site to ensure they meet the Level 1 or 2 design criteria.

<u>3. Construction Inspection</u>. Field inspection is essential to verify that runoff reduction practices are properly installed at the site. This is normally done as a site walk through at the time of final construction inspection. To ensure compliance, communities may want to set the value of the performance bond based on the pre-practice, unadjusted treatment volume for the site to ensure runoff reduction practices are correctly installed.

APPENDIX A: DERIVATION OF RUNOFF REDUCTION RATES FOR SELECT STPs

Runoff reduction (RR) is defined as the average annual reduction in stormwater runoff volume. For stormwater treatment practices, runoff can be reduced via canopy interception, soil infiltration, evaporation, transpiration, rainfall harvesting, engineered infiltration, or extended filtration. Extended filtration includes bioretention or dry swales with underdrains that delay the delivery of stormwater from small sites to the stream system by six hours or more.

Prior to 2003, very few research studies reported flow reductions in the literature, reporting instead on the change in inflow and outflow event mean concentrations (EMC). Recently, more studies have been reporting flow reductions, particularly for LID and STP projects, although data is still limited. Summaries of the runoff reduction performance for individual STPs are discussed in this section.

From a design standpoint, the runoff reduction rates are appropriate for use in the VA spreadsheet up to the water quality storm event. Runoff reduction rates were generally an annual average based on the study site water balance. These rates do not apply to reduction achieved for the channel protection volume storm or larger events. The runoff reduction numbers are dependent on meeting the Level 1 or 2 design criteria and minimum eligibility criteria as set forth in Appendix C and D. Given the limited number of runoff reduction performance studies available, the recommended rates were selected using conservative assumptions, and some of the numbers are considered provisional until more data becomes available.

Green Roofs

Considerable research has been conducted in recent years to define the runoff reduction capability of extensive green roofs (Table A-1). Reported rates for runoff reduction have been shown to be a function of function of media depth, roof slope, annual rainfall and cold season effects. Based on the prevailing climate for the region, a conservative runoff reduction rate for green roofs of 45 to 60% is recommended for initial design.

Table A-1. Volumetric Runoff Reduction by Green Roof				
LID Practice	Location	Runoff Reduction	Reference	
Green Roof	US	40 to 45%	Jarrett et al (2007)	
Green Roof	Germany	54%	Mentens et al (2005)	
Green Roof	MI	30 to 85%	Getter et al (2007	
Green Roof	OR	69%	Hutchinson (2003)	
Green Roof	NC	55 to 63%	Moran and Hunt (2005)	
Green Roof	PA	45%	Denardo et al (2005)	
Green Roof	MI	50 to 60%	VanWoert et al (2005)	
Green Roof	ONT	54 to 76%	Banting et al (2005)	
Green Roof	GA	43 to 60	Carter and Jackson (2007)	
R	R Estimate	45 to 60%		

Rooftop Disconnection

Very limited research has been conducted on the runoff reduction rates for rooftop disconnection, so initial estimates are drawn from research on filter strips, which operate in a similar manner. The research indicates that runoff reduction is a function of soil type, slope, vegetative cover and filtering distance. Table A-2 summarizes filter strip runoff reduction rates within the first 45 feet (where a range is given, the first number is for filtering distance of 5 to 15 ft and the second is from 25 to 45 ft). A conservative runoff reduction rate for rooftop disconnection is 25% for HSG C and D soils and 50% for HSG A and B soils. These values apply to disconnection that meet the feasibility criteria, and do not include any further runoff reduction due to the use of compost amendments along the filter path.

Table A-2. Volumetric Runoff Reduction Achieved by Rooftop Disconnection				
LID Practice	Location	Runoff	Reference	
		Reduction		
Filter Strip	USA	20 to 62	Abu-Zreig et al (2004)	
Filter Strip	USA	40%	Strecker at al (2004)	
Filter Strip	CA	40 to 70	Barrett (2003)	
Runoff Reduction	n Estimate	25 to 50%		

Raintanks and Cisterns

The runoff reduction capability of rain tanks and cisterns has not been extensively monitored, but numerous modeling efforts have assigned a runoff reduction rate. Dual use rain tanks provide indoor potable or grey water and outdoor landscaping irrigation. Modeling research indicates that their runoff reduction capability is limited by tank capacity, and the rate of de-watering between storms, which is strongly influenced by indoor and outdoor water demand, and overflows (Table A-3). The actual rate of runoff reduction for an individual project will require simulation modeling of rainfall and the tank. Based on the prevailing climate for this region, a conservative runoff reduction estimate of 40% is recommended for initial design.

Table A-3. Volumetric Runoff Reduction by Raintanks and Cisterns				
LID Practice Location		Runoff	Reference	
		Reduction		
Dual Use Rain Tanks ¹	AUS (semi-	60 to 90%	Hardy et al (2004)	
	arid)			
Dual Use Rain Tanks	AUS (arid)	40 to 45%	Coombes et al (2002)	
Dual Use Rain Tanks	NZ	35 to 40%	Kettle et al (2004)	
	RR Estimate	40%		

Permeable Pavers

More than a dozen studies are now available to characterize the runoff reduction potential for permeable pavers that are designed with the requisite amount of storage to enable

infiltration beneath the paver. The research studies have been classified into two categories: permeable paver applications that have underdrains and those that do not (Table A-4). Assuming the permeable paver is designed with adequate pretreatment and soil infiltration testing, a conservative runoff reduction rate of 75% is assigned to designs that rely upon full infiltration. Permeable paver applications on HSG C and D soils that typically require underdrains should use the lower runoff reduction rate of 45%.

Table A-4. Volumetric Runoff Reduction by Permeable Pavement					
LID Practice	Location	Runoff Reduction	Reference		
Pervious Pavement *	ONT	99	Van Seters et al (2006)		
Pervious Pavement *	PA	94	Traver et al (2006)		
Pervious Pavement *	FRA	98	Legret and Colandini (1999)		
Pervious Pavement *	NC	100	Bean et al (2007)		
Pervious Pavement *	NC	95 to 98%	Collins et al (2007)		
Pervious Pavement *	WA	97 to 100	Brattebo and Booth (2003)		
Pervious Pavement *	СТ	72	Gilbert and Clausen (2006)		
Pervious Pavement *	UK	78	Jefferies (2004)		
Pervious Pavement #	NC	38 to 66	Collins et al (2007)		
Pervious Pavement #	PA	25-45	Pratt et al (1989)		
Pervious Pavement #	NC	66	Bean et al (2007)		
Pervious Pavement #	UK	53	Jefferies (2004)		
Pervious Pavement #	MD	45 to 60	Schueler et al (1987)		
Pervious Pavement #	Lab	30 to 55	Andersen et al (1989)		
Runoff Reduction Estimate45# to 75*					
* no underdrain collection # underdrain collection					

Grass Channels

Runoff reduction by grass channels is generally low, but is strongly influenced by soil type, slope, vegetative cover, and the length of channel (Table A-5). Recent research indicates that a conservative runoff reduction rate of 10 to 20% can be used depending on whether soils fall in HSG A/B or C/D. The runoff reduction rates can be doubled if the swale is modified to incorporate compost soil amendments.

Table A-5. Volumetric Runoff Reduction Achieved by Grass Channels			
LID Practice	Location	% Runoff	Reference
		Reduction	
Grass Channel	VA	0	Schueler (1983)
Grass Channel	USA	40	Strecker at al (2004)
Grass Channel	NH	0	UNHSC (2007)
Grass Channel	OR	27 to 41	Liptan and Murase (2000)
Runoff Redu	uction Estimate	10 to 20	

Bioretention

More than 10 studies are now available to characterize the runoff reduction rates for bioretention areas. The research can be classified into bioretention applications that possess underdrains and those that do not (and therefore rely on full infiltration into underlying soils) (Table A-6). A conservative runoff reduction rate of 80% is assigned to designs that rely upon full infiltration. Bioretention areas located on HSG C and D soils that typically require underdrains should use the lower runoff reduction rate of 40%.

Table A-6. Volumetric Runoff Reduction Achieved by Bioretention				
LID Practice	Location	% Runoff	Reference	
		Reduction		
Bioretention *	СТ	99%	Dietz and Clausen (2006)	
Bioretention *	PA	86%	Ermilio (2005)	
Bioretention *	FL	98%	Rushton (2002)	
Bioretention *	AUS	73%	Lloyd et al (2002)	
Bioretention #	ONT	40%	Van Seters et al (2006)	
Bioretention #	Model	30%	Perez-Perdini et al (2005)	
Bioretention #	NC	40 to 60%	Smith and Hunt (2007)	
Bioretention #	NC	20 to 29%	Sharkey (2006)	
Bioretention #	NC	52 to 56%	Hunt et al. (2006)	
Bioretention #	NC	20 to 50%	Passeport et al. (2008)	
Bioretention #	MD	52 to 65%	Davis (2008)	
Runoff Reduction Estimate 40# to 80*				
# underdrain design *infiltration design				

Dry Swales

Only a handful of data are available to define the runoff reduction rate for dry swales, but research indicates that they perform as well as, or better than, bioretention with underdrains (Table A-7). Since an underdrain is an integral design feature for dry swales, a conservative runoff reduction of 40% is assigned to dry swales, a value equivalent to that assigned to bioretention with underdrains. If a dry swale lacks an underdrain due to highly permeable soils, or is designed with an underground stone storage layer, the runoff reduction rate can be increased to 60%.

Table A-7. Vo	Table A-7. Volumetric Runoff Reduction Achieved by Dry Swales				
LID Practice	Location	% Runoff	Reference		
		Reduction			
Dry Swale	WA	98%	Horner et al (2003)		
Dry Swale	MD	46 to 54%	Stagge (2006)		
Dry Swale	TX	90%	Barrett et al (1998)		
Bioretention with		20 to 60%	This memo		
underdrains					
Runoff Red	uction Estimate	40 to 60%			

Wet Swales

Limited runoff reduction data is available on wet swales. Wet swales function similarly to wet ponds and wetlands, retaining a permanent pool of water due to intersection with ground water or poorly drained soils. No runoff reduction rate is recommended for wet swales.

Infiltration

The runoff reduction capability of infiltration practices is presumed to be high, given that infiltration is the design intent of the practice. Some surface overflows do occur when the infiltration storage capacity is exceeded. Assuming the practice is designed with adequate pretreatment and soil infiltration testing, a conservative runoff reduction rate of 90% is assigned to infiltration practices. If an underdrain must be utilized, the recommended runoff reduction rate drops to 50% (Table A-8).

Table A-8. Volumetric Runoff Reduction Achieved by Infiltration					
LID Practice	Location	Runoff Reduction	Reference		
Infiltration	NH	90%	UNHSC (2005)		
Infiltration	VA	60%	Schueler (1983)		
Infiltration	PA	90%	Traver et al (2006)		
Infiltration	NC	96-100%	Bright et al (2007)		
Runoff Reduction	n Estimate	50 to 90%			

Extended Detention

In lined extended detention (ED) basins, evaporation reduces a small portion of the runoff volume, and in unlined basins, runoff is further reduced via seepage. Strecker et al. (2004) analyzed the runoff reduction rates for 11 dry extended detention basins in the EPA/ASCE National Stormwater STP Database and found a mean runoff volume reduction of 30%; however, more recent evaluations suggest lower runoff reduction rates (Strecker, personal communication, 2008). Additionally, two ED basins in NC had negligible runoff reduction rates (Hathway et al, 2007e), and a basin in FL sited in very well drained soils had a 70% runoff reduction rate (Harper et al, 1999), which was attributed to groundwater seepage. Based on the prevailing climate for the region, a conservative runoff reduction estimate of 0% for lined basins, and 15% for unlined basins is recommended for initial design.

Soil Amendments

Several studies have examined the effect of soil compost amendments to reduce the volume of runoff produced by lawn runoff from compacted soils (Table A-9). An additional runoff reduction rate of 50% is given when compost amended soils receive runoff from an appropriately designed rooftop disconnection or grass channel. A 75% runoff reduction rate can be used for the runoff from lawn areas that are compost amended, but do not receive any off-site runoff from impervious surfaces.

Table A-9. Volumetric Reduction in Lawn Runoff Due to Compost Amendments				
LID Practice Location Runoff Reference				
		Reduction		
Compost Amendment	WI	74 to 91%	Balusek (2003)	
Compost Amendment	AL	84 to 91%	Pitt et al (1999 and 2005)	
Compost Amendment	WA	29 to 50%	Kolsti et al (1995)	
Compost Amendment	WA	53 to 74%	Hielima (1999)	
Runoff Reduction	on Estimate	50 to 75%		

Filtering Practices, Wetlands, and Wet Ponds

Very little individual performance data is available on the runoff reduction capabilities of sand filters, wet pond, and wetland practices. In pond and wetland applications, evapotranspiration may occur; however, research suggests that the amount of runoff reduced is very low to negligible (Strecker et al, 2004 ; Hathaway et al, 2007a-d). Therefore, a conservative runoff reduction rate of 0% is recommended for filters, wet ponds, and wetlands.

Stormwater Planters, Tree Pits, and Tree Clusters

Only one study has measured the hydrologic capacity of stormwater planters or tree pits to reduce runoff, and it found they had relatively low capability (UNHSC, 2007). The actual runoff reduction capability for these practices is related to their contributing drainage area, runoff storage capacity and rate of overflow or underdrain. Consequently, these practices are assigned a modest runoff reduction capability of 15%. No specific research has been conducted on the runoff reduction rates for tree clusters as set forth in Cappiella et al (2005), although the value of trees in reducing runoff has been established by Portland BES (2003) and PA DEP (2006). These manuals assign a runoff reduction rate of 6 cubic feet per qualifying deciduous tree and 10 cubic feet per evergreen tree. If planting bed is compost amended, or tree cluster is designed to accept off-site runoff, a higher rate of runoff reduction may be used.

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APPENDIX B: DERIVATION OF EMC POLLUTANT REMOVAL RATES FOR SELECT STPs

Pollutant removal efficiency refers to the pollutant reduction from the inflow to the outflow of a system. Pollutant removal efficiency can be calculated using variety of computations, but the two most common methods are event mean concentration (EMC) efficiency and mass or load efficiency. EMC efficiency is derived by averaging the influent and effluent concentrations for storm events, and then calculating the median change. Mass efficiency is calculated by determining the pollutant load reduction from the influent to effluent, and is influenced by the volume of water reduced by the practice (e.g., runoff reduction).

Depending on the method used, reported removal efficiencies of stormwater best management practices (STPs) can vary widely and are often inconsistent. Further, removal efficiencies do not adequately address runoff volume reductions in STPs (Strecker et al, 2004; Jones et al, 2008). However, for the purposes of this method, reporting EMC based pollutant removal efficiencies can isolate key STP pollutant removal mechanisms and offers a supplemental approach to increase STP performance apart beyond runoff reduction.

The following sections discuss the derivation of EMC based pollutant removal efficiencies of STPs. The NPRPD (CWP, 2000) details the pollutant removal efficiencies of several STPs that were derived using several different methods. Studies reporting EMC pollutant removal in the NPRPD were isolated and included in the analysis. Further, EMC pollutant removal numbers were compiled from recent studies. When possible, a median and 75th percentile value for nutrient PR was determined.

It should be noted that the data used to estimate pollutant removal was derived from practices in good conditions; most studies focused on STPs that were constructed within three years of monitoring. Further, the actual EMC pollutant removal performance can be strongly influenced by the influent quality. Since pollutant removal rates are usually dependent on site characteristics and STP geometry, the EMC based pollutant removal numbers are dependent on meeting the level 1 or 2 design criteria (Appendix C) or the eligibility criteria for ESD (Appendix D). Due to the limited number of performance studies, conservative EMC pollutant removal rates were selected. In several cases, provisional numbers are set forth until more data becomes available.

Green Roofs

In recent years, several studies have been conducted on the nutrient removal capabilities of green roofs. Results confirm that green roofs initially leach nutrients from the compost contained the growth media used to support initial plant growth (Table B-1). Several studies have suggested that the leaching may subside over time; however, the extent to which nutrient leaching decreases has not been quantified. Media with high

compost content will leach more nutrients than media with lower compost content. Therefore, to minimize the export of nutrients, media should be selected with the lowest compost content to support the growth of the desired roof vegetation. No pollutant removal credit for nitrogen or phosphorus is recommended.

Table B-1. Pollutant Removal Achieved by Green Roofs					
LID Practice:	Location	ТР	TN	Study	
Green Roof ¹		Removal	Removal		
Green Roof	NC	negative	Negative	Moran et al, 2005	
Green Roof	OR	negative	Negative	Hutchinson, 2003	
Green Roof	CAN	negative	Negative	Banting et al, 2005	
EMC PR estimate 0% 0%					
¹ Pollutant removal values are EMC based for all studies					
⁺ Study included in NPRPD (CWP, 2007)					

Disconnection (Vegetated Filter Strips)

Limited research has been conducted on the pollutant removal rates for rooftop disconnection, so initial estimates are drawn from research on filter strips, which operate in a similar manner. The research indicates that nutrient reduction is a function of filtering distance and vegetative cover (Abu-Zreig et al, 2003; Barrett et al, 1998; CALTRANS, 2004; Goel et al, 2004). Since very little information regarding the EMC based nutrient removal rates of vegetated filter strips has been published, no pollutant removal rate for TP and TN is recommended at this initial stage.

Raintanks and Cisterns

Limited research has been conducted to evaluate the pollutant removal capabilities of rain tanks and cisterns, however, it is generally understood that no primary pollutant removal benefits exist (MPAC, ND). Based on this assumption, no pollutant removal credit for nitrogen or phosphorus is recommended for raintanks and cisterns.

Permeable Pavement

While several studies have documented high heavy metal and TSS removal efficiencies of permeable pavements, few studies have evaluated permeable pavement nutrient removal capabilities. Limited results indicate that permeable pavement TP and TN removal rates vary widely (Table B-2). TP can potentially be reduced by adsorption to the aggregate and soils in the pavement subbase layers, but may also leach from underlying soils or surface fill material in pavement void spaces. Provisional EMC pollutant removal rates of 25% for both total phosphorus and total nitrogen are recommended for initial design.

Table B-2. Pollutant Removal Achieved by Permeable Pavements					
LID Practice:	Location	Pollutant	Pollutant	Study	
Permeable Pavement ¹		Removal	Removal		
		(TP)	(TN)		
Permeable Pavement#	Lab	60%		Day et al, 1981	
Permeable Pavement#	CAN	0%		James and Shahin, 1998	
Permeable Pavement#	GA	10%	negative	Dreelin et al, 2006	
Permeable Pavement#	NC	65%	36%	Bean et al, 2007 ⁺	
Permeable Pavement#	NC	negative	negative	Bean, 2005 ⁺	
Permeable Pavement#	NH	38%		UNH, 2007	
Permeable Pavement#	NC	0%	25%*	Collins et al., 2008	
Permeable Pavement#	СТ	34%	88%	Gilbert and Clausen, 2006	
EMC PR estimate 25% 25%					
¹ Pollutant removal values are EMC based for all studies					
⁺ Study included in NPRPD (CWP, 2007)					

* for one pavement type only

underdrain design

Grass Channels (Drainage Swales)

Several studies have documented the nutrient removal rates of drainage swales (Table B-3). Nutrient removal is generally low, but is influenced by vegetative cover and flow velocity. The removal of mowed grass clippings may also increase nutrient removal. Fertilization of channel vegetation should be avoided. Conservative pollutant removal rates of 15% for TP and 20% for TN are recommended for initial design.

Table B-3. Pollutant Removal Achieved by Grass Channels					
LID Practice:	Location	Pollutant	Pollutant	Study	
Drainage Swale ¹		Removal	Removal		
		(TP)	(TN)		
Grass Channel	MD	0%	37%	OWML, 1983 ⁺	
Grass Channel	MD	0%	negative	OWML, 1983 ⁺	
Grass Channel	ΤХ	34 to 44%	38%	Walsh et al, 1995 $^+$	
				Welborn and Veehuis,	
Grass Channel	TX	negative	negative	1987 +	
Grass Channel	FL	13%	21%	Harper, 1988 ⁺	
Grass Channel	FL	25%	11%	Yousef et al, 1986^+	
Grass Channel	WA	29-45		Seattle Metro, 1992 ⁺	
Grass Channel	CA	negative	30%	CALTRANS, 2004	
				Schueler and Holland,	
Grass Channel	USA	29		2000 (article 116)	
EMC PR estimate 15% 20%					
¹ Pollutant removal values are EMC based for all studies except NPRPD					
⁺ Study included in NPRPD (CWP, 2007)					

Bioretention

Several recent studies have indicated that bioretention practices are effective at removing nutrients, as well as metals, pathogens, oil and grease. Much of this research has reported mass based pollutant removal rates, but ten studies reporting EMC based removal rates are now available (Table B-4). The extent of TP removal is related to bioretention cell depth, mulching, plant cover, and the organic matter content of the soil media. The primary phosphorus removal mechanism is soil adsorption. It is imperative that the P-index of the media be tested to ensure a low number (less than 30), as earlier studies have found that soil media with a high P-index will leach phosphorus.

Nitrogen is removed through mineralization and denitrification near the surface of bioretention cells and also by denitrification in anaerobic zones that often develop deeper in the cells. Design of an internal water storage zone (sump) using an upturned underdrain may increase TN removal. A summary of bioretention mass removal included in the NPRPD lists lower median and 75th percentile pollutant removal rates for TP; however, many of these earlier studies tested practices with high P-index media. Conservative EMC pollutant removal rates of 25 to 50% for TP removal and 40 to 60% for TN removal are recommended for initial design. TP removal is credited only if the media is tested to ensure that the media P-index is less than 30.

Table B-4. Pollutant Removal Achieved by Bioretention					
LID Practice: Bioretention ¹	Location	Pollutant Removal (TP)	Pollutant Removal (TN)	Study	
NPRPD (N=10)		5 ^a -30 ^b	46 ^a -55 ^b	CWP, 2007	
Bioretention#	MD	81%		Davis et al., 2001	
Bioretention#	MD	65%	49%	Davis et al., 2006	
Bioretention#	MD	87%	59%	Davis et al., 2006	
Bioretention#	Lab	81%	60%	Davis et al., 2006	
Bioretention#	PA	1%	48%	Ermilio, 2005 ⁺	
Bioretention#	NC	8%	61%	Smith and Hunt, 2006^+	
Bioretention#	NC	32%	38%	Hunt et al. 2008	
Bioretention#	NC	60%	54%	Passeport et al. 2008	
Bioretention#	NC	66%	62%	Sharkey, 2006	
Bioretention#	VA	13%		Yu and Stopinski, 2001 ⁺	
EMC PR estimate 25 to 50% 40 to 60%					
¹ Pollutant removal values are EMC based for all studies except NPRPD					
^a Median pollutant removal rate ^b 75 th Percentile pollutant removal rate					
⁺ Study included in NPRPD (CWP, 2007)					

underdrain design

Water Quality Swales

Less monitoring data is available to define the EMC pollutant removal rate for water quality swales, which include wet swales and dry swales with an underdrain. Research suggests that pollutant removal mechanisms of dry swales are similar to those of a bioretention cell with an underdrain, because a portion of water is filtered through a soil media. Wet swales, which typically contain a shallow permanent pool, may function similar to, but less efficient than, wetlands or wet ponds with respect to pollutant removal. Conservative and provisional EMC pollutant removal rates of 20 to 40% for TP and 25 to 35% for TN are recommended for the initial design of both wet and dry swales (Table B-5).

Table B-5. Pollutant Removal Achieved by Water Quality Swales					
LID Practice:	Location	Pollutant	Pollutant	Study	
Water Quality		Removal	Removal		
Swales ¹		(TP)	(TN)		
Wet swale	FL	17%	40%	Harper, 1988 ⁺	
Wet swale	WA	39		Koon, 1995 ⁺	
Dry swale	AUS	65%	52%	Fletcher et al, 2002	
Dry swale with					
Underdrain	ΤХ	31		Barrett et al, 1997	
Wet Ponds		50 to 75%	30 to 40%	This study	
Bioretention with					
Underdrain		25 to 50%	25%	This study	
EMC PR estimate 20 to 40% 25 to 35%					
¹ Pollutant removal values are EMC based for all studies					
⁺ Study included in NPRPD (CWP, 2007)					

Infiltration

Because of the difficulty associated with monitoring infiltration practices, very limited data is available on EMC nutrient removal capability. Studies have indicated that stormwater pollutants, including nutrients, can be filtered out in the soils underlying infiltration basins (Mikkelson et al, 1994; Barraud et al, 1999; Dechesne et al, 2003). A summary of 12 infiltration practices included in the NPRPD lists the median and 75th percentile mass pollutant removal rates as 65 to 96 for total phosphorus (TP), and 42 to 65 for total nitrogen (TN). However, the majority of mass removal in infiltration practices occurs in the form of runoff reduction (Appendix A), so provisional EMC pollutant removal rates of 25% for TP removal and 15% for TN removal are specified until more research becomes available.

Extended Detention

Extensive research on ED ponds has indicated that these practices can effectively remove particulate pollutants, primarily thorough sedimentation. Documented nutrient removal rates are variable (Table B-6). Based on several studies, conservative EMC pollutant

removal rates of 15% for TP and 10% for TN are recommended. The EMC pollutant removal differs from the removal rates in the NPRPD, which did not include any ED studies that analyzed EMC based pollutant removal.

Table B-6. Pollutant Removal Achieved by Extended Detention				
Practice:	Location	Pollutant	Pollutant	Study
Extended Detention ¹		Removal	Removal	
		(TP)	(TN)	
NPRPD (N=10)		$20^{a}-25^{b}$	24 ^a -31 ^b	CWP, 2007
		15 to		
Dry ED pond	CA	39%	14%	CALTRANS, 2004
Dry ED pond	NC	0%	10 to 13%	Hathaway et al, 2007e,f
Dry ED pond	NJ	34%	0%	Harper et al, 1999 ⁺
				Middleton and Barrett,
Dry ED pond	ΤХ	7%		2006
EMC PR estimate 15% 10%				
¹ Pollutant removal values are EMC based for all studies except NPRPD				
^a Median pollutant removal rate				
^b 75 th Percentile pollutant removal rate				

⁺ Study included in NPRPD (CWP, 2007)

Soil Amendments

Few studies have reported on the pollutant removal capabilities of amended soils. Both Glanville, et al. (2003) and Pitt et al, (2005) found that the pollutant concentrations in runoff from compost amended soils were higher than in runoff from un-amended soils. Pitt et. al. (2005) found that subsurface flows had an increased amount of nitrogen and phosphorus as compared to un-amended soils. This difference was present at newly constructed sites but was less prominent at older sites. Due to the high compost or organic matter content that is added to amended soils, it can be assumed that negligible removal of nutrients would occur, and nutrients may, in fact, leach from soil runoff, similar to documented pollutant removal of green roof media containing compost. As such, no pollutant removal credit for nitrogen or phosphorus is recommended for soil amendments.

Filtration

Numerous studies have evaluated the nutrient removal capabilities of various stormwater filtration practices (Table B-7). Phosphorus is removed via chemical precipitation in the filter bed media, and although filters may export nitrates, studies have indicated that TN is typically reduced. The use of some organic materials in the filter bed, which can improve heavy metal removal rates, may cause nutrient leaching. An analysis of individual studies in which the EMC pollutant removal rates were reported yielded EMC removal rates of for TP (N=7 studies) and TN (N=4 studies) similar to the pollutant removal rates in the NPRPD (N=18 studies). Since runoff reduction in filtration practices is negligible (Appendix A), mass removal and EMC removal rates are roughly

equivalent. Since so few studies report EMC removal rates, filtration practices were assigned based on their NPRPD removal rates of 60 to 65% for TP, and 30 to 45% for TN.

Table B-7. Pollutant Removal Achieved by Filtration						
Practice:	Location	ТР	TN	Study		
Sand Filters ¹		Removal	Removal			
NPRPD (N=18)		59 ^a -66 ^b	32 ^a -47 ^b	CWP, 2007		
Sand Filter	TX	39 %	22%	Barrett, 2003		
Sand Filter	VA	66%	47%	Bell et al, 1995^+		
Peat Sand Filter	TX	48%	30 to 51%	LCRA, 1997 ⁺		
Sand Filter	WA	20 to 41%		Horner, 1995 ⁺		
Sand Filter	TX	45%	15%	Barton Springs, 1996 ⁺		
Organic filter	WI	88%		Corsi and Greb, 1997 ⁺		
Compost filter						
EMC PR estimate 60 to 65% 30 to 45%						
¹ Pollutant removal values are EMC based for all studies except NPRPD						
^a Median pollutant removal rate						
^b 75 th Percentile pollutant removal rate						

⁺ Study included in NPRPD (CWP, 2007)

Wetlands

Studies indicate that wetlands can effectively remove TP and TN, primarily through sedimentation and plant nutrient uptake (Table B-8). Nutrient removal is related to the vegetative covering, wetland geometry, and the drawdown time of the temporary storage volume. An analysis of individual studies in which the EMC pollutant removal rates were reported yielded EMC removal rates of for TP (N=8 studies) and TN (N=4 studies) similar to the pollutant removal rates in the NPRPD (N=40 studies). Since runoff reduction in wetland practices is negligible (Appendix A), mass removal and EMC removal rates are roughly equivalent. Wetlands were therefore assigned EMC pollutant removal rates in the NPRPD: 50 to 75% for TP, and 25 to 55% for TN.

Wet Ponds

Numerous studies have evaluated the nutrient removal capabilities of wet ponds (Table B-9). Several factors appear to affect removal rates, such as the treatment volume captured, presence of emergent vegetation, and length of the flow path in the pond. The establishment of a diverse, dense plant community around the perimeter of the pond may increase nutrient removal, and may also discourage water fowl activity, potentially reducing organic nutrient and pathogen inputs. An analysis of individual studies in which the EMC pollutant removal rates were reported yielded EMC removal rates of for TP (N=16 studies) and TN (N=12 studies) similar to the pollutant removal rates in the NPRPD (N=46 studies). Since runoff reduction in wet pond practices is negligible (Appendix A), mass removal and EMC removal rates were considered to be equivalent. Wet ponds were therefore assigned pollutant removal rates based on the values in the NPRPD: 50 to 75% for TP, and 30 to 40% for TN.

Table B-8. Pollutant Removal Achieved by Wetlands					
Practice:	Location	ТР	TN	Study	
Wetlands ¹		Removal	Removal		
NPRPD (N=40)		48^{a} -76 ^b	24 ^a -55 ^b	CWP, 2007	
Wetland	FL	28%	10%	Martin, 1988^+	
Wetland	FL	48%	13%	Blackburn et al, 1986 ⁺	
Wetland	WA	33%		Koon, 1995 ⁺	
Wetland	FL	57%		Rushton and Dye, 1993 ⁺	
Wetland	VA	69%		Yu et al, 1998 ⁺	
Wetland	VA	15%		Yu et al, 1998 ⁺	
Submerged gravel					
wetland	CA	46%	Negative	Reuter et al, 1992 ⁺	
Wetland	NC	45%	35 to 45%	Hathaway et al, 2007a,b	
EMC PR estimate 50 to 75% 25 to 55%					
¹ Pollutant removal values are EMC based for all studies except NPRPD					
^a Median pollutant removal rate					

^a Median pollutant removal rate ^b 75th Percentile pollutant removal rate ⁺ Study included in NPRPD (CWP, 2007)

Table B-9. Pollutant Removal Achieved by Wet Ponds						
Practice:	Location	ТР	TN	Study		
Wet Ponds ¹		Removal	Removal			
NPRPD (N=46)		$52^{a}-76^{b}$	31 ^a -41 ^b	CWP, 2007		
Wet Pond	TX	87%	50%	City of Austin, TX 1996 ⁺		
Wet Pond	WA	19%		Comings et al, N.D ⁺		
Wet Pond	FL	55%	12%	Cullum, 1984 ⁺		
Wet Pond	FL	30%	16%	Gain, 1996 ⁺		
Wet Pond	FL	40%		Kantrowitz & Woodham, 995 ⁺		
Wet Pond	FL	22%	15%	Martin, 1988 ⁺		
Wet Pond	CAN	72%		SWAMP, 2000 ⁺		
Wet Pond	CA	29%	0%	Taylor et al, 2001		
Wet Pond	NC	57%	40%	Mallin et al, 2002		
Wet Pond	CA	5%	51%	CALTRANS, 2004		
Wet Pond	NC	15 to 41%	19 to 23%	Hathaway et al, 2007c,d		
Wet ED pond	CAN	37%	28%	Fellows et al, 1999 ⁺		
Wet ED pond	CO	52%	55%	LCRA, 1997 ⁺		
Wet ED pond	FL	75%	28%	Rushton et al, 1995 ⁺		
Wet ED pond	FL	50%	25%	Rushton et al, 2002 ⁺		
Wet ED pond	CAN	56 to 65%		SWAMP, 2000		
EMC PR estimate 50 to 75% 30 to 40%						
¹ Pollutant removal values are EMC based for all studies except NPRPD						

^a Median pollutant removal rate ^b 75th Percentile pollutant removal rate ⁺ Study included in NPRPD (CWP, 2007)

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APPENDIX C: LEVEL 1 AND 2 STP DESIGN FACTORS

The following tables assign design factors to Level 1 or 2 that will achieve the indicated average runoff reduction and phosphorus removal rates.

- C-1 Green Roof
- C-2 Permeable Pavers
- C-3 Bioretention
- C-4 Dry Swale
- C-5 Wet Swale
- C-6 Infiltration
- C-7 Extended Detention Pond
- C-8 Filtering Practice
- C-9 Constructed Wetland
- C-10 Wet Pond

The base pollutant removal and runoff reduction are the median values for Level 1, whereas Level 2 is the 75th percentile values. These tables do not include the standard setbacks, restrictions, feasibility constraints and minimum design features that apply to each practice at all site applications.

Table C-1. Green Roof Design Guidance	
Level 1 Design (RR:45; TP:0; TN:0)	Level 2 Design (RR: 60; TP:0; TN:0)
Depth of media four to six inches ¹	Media depth greater than six inches
Soil media not tested for P-index	Soil media with P index less than 10
Green roof receives roof runoff	Green roof does not receive roof runoff or
	is designed with additional media depth
All Designs: shall be in conformance to ASTM (2005) International Green Roof	
Standards. Appropriate media and plant selection for harsh rooftop conditions and	

shallow media depths. Filter media mix should have the minimum organic matter/nutrient content to maintain fertility for plant growth but not contribute to nutrient leaching.

¹If media depth is less than 4 inches, the runoff reduction credit is adjusted that each inch of media provides a 10% reduction in runoff volume.

Table C-2. Permeable Paver Design Guidance	
Level 1 Design (RR:45; TP:25; TN:25)	Level 2 Design (RR: 75 TP:25; TN:25)
TV = (1.0)(Rv)(A)	TV = (1.1)(Rv) (A)
Soil infiltration less than one-inch/hr	Soil infiltration rate exceeds one-inch/hr
Underdrain needed	Underdrain not required
CDA exceeds the pervious paver area	CDA = The pervious paver area
Slopes from 2 to 5%	Slopes less than 2%

Table C-3. Bioretention Design Guidelines	
Level 1 Design (RR:40; TP:25; TN:40)	Level 2 Design (RR:80; TP:50; TN:60)
TV = (1.0)(Rv)(A)	TV = (1.25) (Rv)(A)
SA of filter exceeds 3% of CDA	SA of filter bed exceeds 5% of CDA
Filter media at least 24" deep	Filter media at least 36" deep
One form of accepted pretreatment	Two or more forms of accepted pretreatment
At least 75% plant cover w/ mulch	At least 90% plant cover, including trees.
One cell design	Two cell design
Underdrain needed	Infiltration design or underground stone sump
All Designs: acceptable media mix tested	d for phosphorus index, does not treat
stormwater hotspot or baseflow.	

Table C-4. Dry Swale Design Guidance	
Level 1 Design (RR:40; TP:20; TN:25)	Level 2 Design (RR:60; TP:40; TN: 35)
TV = (1.0)(Rv)(A)	TV = (1.1)(Rv)(A)
Swale slopes from <0.5% or >2.0%	Swale slopes from 0.5% to 2.0%
Soil infiltration rates less than 0.5 in	Soil infiltration rates exceed one inch
Swale served by underdrain	Lacks underdrain or uses underground stone sump
On-line design	Off-line or multiple treatment cells
Media depth less than 18 inches	Media depth more than 24 inches
Turf cover	Turf cover, with trees and shrubs
All Designs: acceptable media mix tested	l for phosphorus index

Table C-5. Wet Swale Design Guidance	
Level 1 Design (RR:0; TP:20; TN:25)	Level 2 Design (RR:0; TP:40; TN:35)
TV = (1.0)(Rv)(A)	TV = (1.25)(Rv)(A)
Swale slopes more than 1%	Swale slopes less than 1%
On-line design	Off-line swale cells
No planting	Wetland planting within swale cells
Turf cover in buffer	Planting trees/shrubs within swale cells
Note: Generally recommended only for flat coastal plain conditions with high water	
table. Linear wetland always preferred to we	et swale

Table C-6. Infiltration Design Guidelines	
Level 1 Design (RR:50; TP:25; TN:15)	Level 2 Design (RR:90; TP:25; TN:15)
TV = (1.0)(Rv)(A)	TV = (1.1)(Rv)(A)
Maximum CDA of one acre	Max CDA of 0.5 acre, nearly 100% IC
At least one form of pretreatment	At least two forms of pretreatment
Soil infiltration rate of 0.5 to 1.0 in/hr	Soil infiltration rates of 1.0 to 4.0 in/hr
Underdrain needed due to soils	No underdrain utilized
All Designs: no hotspot runoff	

Table C-7. Extended Detention (ED) Pond Guidance	
Level 1 Design (RR:0; TP:15; TN:10)	Level 2 Design (RR:15; TP:15; TN:10)
TV = (1.0)(Rv)(A)	TV = (1.25)(Rv) (A)
At least 15% of TV in permanent pool	More than 40% of TV in deep pool or wetlands
Flow path at least 1:1	Flow path at least 1:5 to 1
Average ED time of 24 hours or less	Average ED time of 36 hours
vertical ED fluctuation exceeds 4 feet	Maximum vertical ED limit of 4 feet
Turf Cover on floor	Trees and wetlands in the planting plan
Forebay and micropool	Additional cells or treatment methods (e.g.,
	sand filter or biotretention on pond floor)
CDA less than ten acres	CDA greater than ten acres

Table C-8. Filtering STP Design Guidance	
Level 1 Design (RR:0; TP:60; TN:30)	Level 2 Design (RR:0 ¹ ; TP:65; TN:45)
TV = (1.0)(Rv)(A)	TV = (1.25)(Rv)(A)
One cell design	Two cell design
Sand media	Sand media w/ organic layer
CDA contains pervious area	CDA is nearly 100% impervious
Not a confirmed stormwater hotspot	Site is a confirmed stormwater hotspot
¹ can be increased to up to 50% if or seco	ond cell is used for infiltration

Table C-9. Constructed Wetland Design Guidance	
Level 1 Design (RR:0; TP:50; TN:25)	Level 2 Design (RR:0; TP:75; TN:55)
TV = (1.0)(Rv)(A)	TV = (1.5)(Rv)(A)
Single cell (with forebay)	Multiple cells or pond/wetland design
ED wetland	No ED in wetland
Uniform wetland depth	Diverse microtopography
Mean wetland depth more than one foot	Mean wetland depth less than one foot
Wetland SA/CDA ratio less than 3%	Wetland SA/CDA ratio more than 3%
Flow path 1:1 or less	Flow path 1.5:1 or more
Emergent wetland design	Wooded wetland design

Table C-10 Wet Pond Design Guidance	
Level 1 Design (RR:0; TP:50; TN:30)	Level 2 Design (RR:0; TP:75; TN:40)
TV = (1.0)(Rv)(A)	TV = (1.5)(Rv) (A)
Single Pond Cell (w/ forebay)	Wet ED or Multiple Cell Design
Pool Depth Range of 3 to 12 feet	Pool Depth Range of 4 to 8 feet
Flow path 1:1 or less	Flow path 1.5:1 or more
Pond intersects with groundwater	Adequate water balance
CDA less than 15 acres	CDA greater than 15 acres

APPENDIX D: MINIMUM CRITERIA FOR SELECT ESD PRACTICES

From a design standpoint, it is still important to establish qualifying criteria for the following ESD practices:

- Site Reforestation
- Soil Restoration
- Sheetflow to Conserved Open Space
- Rooftop Disconnection
- Grass Channels

The updated design criteria for these ESD practices are provided in the tables below. In most cases, the design criteria were based on the original qualifying credit criteria contained in the 2000 MDE Manual, but they have been updated to reflect local experience, further credit details in other manuals produced since 2000 (e.g., Minnesota, Credit River, DCR). The soil restoration and site reforestation criteria were drafted using recent research.

Table D-1. Site Reforestation

Description: Site reforestation involves planting trees on existing turf or barren ground at a development site with the explicit goal of establishing a mature forest canopy that will intercept rainfall, increase evapo-transpiration and enhance soil infiltration rates. Reforestation areas at larger development sites and for individual trees for smaller development sites are eligible under certain qualifying conditions.

Computation: A runoff coefficient of twice the forest runoff coefficient may be used for the entire combined areas of reforestation in the contributing drainage area, since it may take several decades for the replanted area to mature and provide full hydrologic benefits. If reforestation is combined with soil amendments, then the forest cover coefficient area can be used instead. The runoff reduction calculation for individual qualifying trees or tree clusters is 6 cubic feet per deciduous tree and 10 cubic feet per evergreen tree ¹

Eligibility for Reforestation Practice (sites greater than one acre in size)

- The minimum contiguous area of reforestation must be greater than 5000 square feet
- A long term vegetation management plan must be prepared and filed with the local review authority to maintain the reforestation area in a natural forest condition
- The reforestation area must be protected by a perpetual stormwater easement or deed restriction that indicates that no future development or disturbance can occur within the area
- Reforestation methods must achieve 75% forest canopy within ten years
- The planting plan must be approved by the appropriate local forestry or conservation authority, including any special site preparation needs
- The construction contract should contain a care and replacement warranty extending at least three growing seasons to ensure adequate growth and survival of the plant community
- The reforestation area shall be shown on all construction drawings and ESC plans during construction

Eligibility for Individual Tree Practice (Sites less than one acre in size).

• Qualifying trees on small sites include native tree at less two inches in caliper planted in expanded tree pits with adequate soil volume to ensure future growth and survival

¹ The individual tree runoff credits were developed from data contained in Portland BES(2004), PA DEP (2006) and Cappiella et al (2005a and 2005b)

Table D-2. Soil Restoration Criteria

Application: Compost amended soils can be used to reduce the generation of runoff from compacted urban lawns and to enhance the runoff reduction performance of downspout disconnections and grass channels (**Note: See Draft Soil Restoration Specification**).

Computation: A runoff reduction rate of 50% is given when compost amended soils receive runoff from an appropriately designed rooftop disconnection or grass channel. A 75% runoff reduction rate can be used for the runoff from lawn areas that are compost amended, but do not receive any off-site runoff from impervious surfaces.

Suitability for Soil Restoration: Compost amended soils are suitable for any pervious area where soils have been or will be compacted by the grading and construction process. They are particularly well suited when existing soils have low infiltration rates (HSG C and D) and when the pervious area will be used to filter runoff (downspout disconnections and grass channels). The area or strip of amended soils should be hydraulically connected to the stormwater conveyance system. Compost amendments are not recommended where:

- Existing soils have high infiltration rates (i.e., A HSG)
- The water table or bedrock is located within 1.5 feet of the soil surface.
- Slopes exceed ten percent
- Existing soils are saturated or seasonally wet
- They would harm roots of existing trees (stay outside the tree drip line)
- The downhill slope runs toward an existing or proposed building foundation

Sizing: Several simple sizing criteria are used when soil compost amendments are used to enhance the performance of a downspout disconnection

- Flow from the downspout should be spread over a 10 foot wide strip extending down-gradient from the building to the street or conveyance system.
- Existing soils in the strip will be scarified or tilled to a depth of 12 to 18 inches and amended with well-aged compost to achieve a organic matter content in the range of 8 to 13%.
- The depth of compost amendment is based on the relationship of the contributing rooftop area to the area of the soil amendment strip, using the guidance presented in the Bay-wide soil restoration design specification

Similar sizing criteria are used when soil compost amendments are used to enhance the performance of a grass channel

- Flow in the grass channel should be spread over a 10 foot wide strip extending the length of the bottom of the channel
- Existing soils in the strip will be scarified or tilled to a depth of 12 inches and soils mixed with 6 to 8 inches of well-aged compost to achieve an organic matter content in the range of 8 to 13%.
- The amended area will need to be rapidly stabilized with perennial, salt tolerant grass species. For grass channels on steep slopes, it may be necessary to install a

Table D-2. Soil Restoration Criteria

protective biodegradable geotextile fabric

• Designers will need to ensure that the final elevation of the grass channel meets original hydraulic capacity

Design Specifications: Leaf compost should be made exclusively of fallen deciduous leaves with less than 5% dry weight of woody or green yard debris materials. The compost shall contain less than 0.5% foreign material such as glass or plastic contaminants and be certified as pesticide free. The use of leaf mulch, composted mixed yard debris, biosolids, mushroom compost or composted animal manures is prohibited. The compost shall be matured and been composted for a period of at least one year and exhibit no further composition. Visual appearance of leaf matter in the compost is not acceptable. The compost should have a dry bulk density ranging from 40 to 50 lbs/ft³, a pH between 6 to 8 and a CEC in excess 50 meq/100 grams dry weight.

Construction Sequence: The construction sequence for compost amendments differs depending whether the practice will be applied to a large area or a narrow filter strip such as in a rooftop disconnection or grass channel. For larger areas, a typical construction sequence is as follows.

- 1. Prior to building, the proposed area should be deep tilled to a depth of 2 to 3 feet using a tractor with two deep shanks (curved metal bars) to create rips perpendicular to the direction of flow.
- 2. A second deep tilling is needed after final building lots have been graded to a depth 12 to 18 inches
- 3. An acceptable compost mix is then incorporated into the soil using a rototiller or similar equipment at the volumetric rate of one part compost to two parts soils
- 4. The site should be leveled and seed or sod used to establish a vigorous grass cover. Lime or irrigation may initially be needed during start
- 5. Compost amendment areas exceeding 2500 square feet should employ simple erosion control measures, such as silt fence, to reduce the potential for erosion
- 6. If the soil restoration area will receive any upslope runoff, then erosion control measures are needed to keep sediments from upslope runoff from compromising the amended area, particularly during construction
- 7. Construction inspection involves digging a test pit to verify the depth of mulch, amended soil and scarification. A rod penetrometer should be used to establish the depth of uncompacted soil at one location per 10,000 square feet

The first step is usually omitted when compost is used for narrower filter strips.

Table D-3. Sheetflow To Conserved Open Space

Description: Sending sheetflow from developed areas of the site to protected conservation areas

Computation: The runoff coefficient for conservation area will be forest or restoration area, depending on predevelopment land cover. Qualifying contributing areas include any turf and impervious cover that is hydrologically connected to the protected conservation area and is effectively treated by it. A 75% runoff reduction practice is given for qualifying HSG A and B soils, and a 50% runoff reduction is given for qualifying HSG C and D soils.

Basic Eligibility for the Conservation Area

- The minimum combined area of all natural areas conserved within the appropriate drainage area must exceed 0.5 acres
- No major disturbance may occur within the open space during or after construction (i.e., no clearing or grading allowed except temporary disturbances associated with incidental utility construction, restoration operations or management of nuisance vegetation). The conservation area shall not be stripped of topsoil. Some light grading may be needed at the boundary using tracked vehicles to prevent compaction
- The limits of disturbance should be clearly shown on all construction drawings and protected by acceptable signage and silt fencing
- A long term vegetation management plan must be prepared to maintain the conservation area in a natural vegetative condition. Managed turf is not considered an acceptable form of vegetative management, and only the passive recreation areas of dedicated parkland are eligible for the practice (e.g., ball fields and golf courses are not eligible)
- The conservation area must be protected by a perpetual easement or deed restriction that assigns the responsible party to ensure no future development, disturbance or clearing can occur within the area.
- The practice does <u>not</u> apply to jurisdictional wetlands that are sensitive to increased inputs of stormwater runoff

Basic Eligibility for the Runoff Generating Area

- The maximum contributing sheet flow path from adjacent pervious areas should not exceed 150 feet
- The maximum contributing sheet flow path from adjacent impervious areas should not exceed 75 feet
- If the contributing flow path has a slope greater than 3%, graded terraces should be placed every 20 feet along the flow path
- Runoff should enter the boundary of the open space as sheetflow for the one-inch storm. A depression, berm or level spreader may be used to spread out concentrated flows generated during larger storm events.

Table D-4. Rooftop Disconnection

Description:

This runoff reduction practice is offered when rooftop runoff is disconnected, and then filtered, treated, or reused before it moves from roof to the storm drain system.

Computation:

Two kinds of practices are allowed. One is for simple rooftop disconnection, whereas the second involves disconnection combined with supplementary runoff treatment involving:

(a) Compost amended soils in the filter path

(b) Installation of rain gardens or dry wells

(c) Storage and reuse in a rain tank, cistern or foundation planter.

Simple disconnection is assigned a runoff reduction rate of 50% on A/B soils (25% on C/D soils). Disconnection to amended soils is assigned a 50% reduction. Disconnection to rain gardens or dry wells is assigned a 75% reduction on A/B soils (50% for C/D soils). The runoff reduction for rain tanks and cisterns is 40%, but varies depending on design and the degree of water reuse. See Figure D-1 to determine the most appropriate rooftop disconnection option.

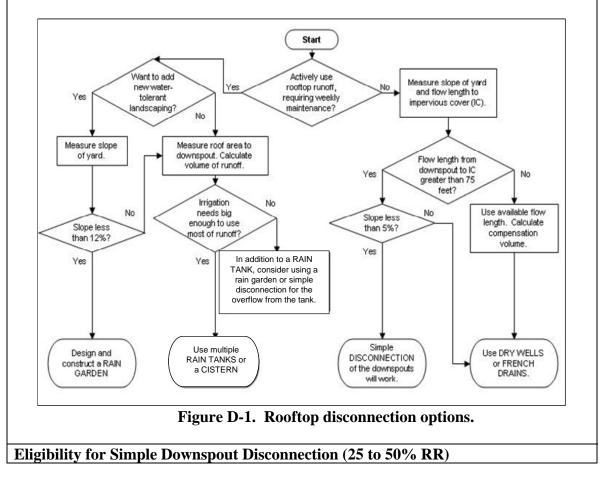


Table D-4. Rooftop Disconnection

- Simple disconnection is only allowed for residential lots greater than 6000 sf. For lot sizes smaller than 6000 sf, disconnection with supplementary runoff treatment can be considered.
- The contributing flow path from impervious areas should not exceed 75 feet
- The disconnection length must exceed the contributing flow path
- If suitable soil amendments are provided (see Table E-2), the 50% runoff reduction rate may be used for C/D soils
- A compensatory mechanism is needed if the disconnection length is less than 40 feet and/or the Hydrologic Soil Group is in the C or D Category.
- Pervious areas used for disconnection should be graded to have a slope in the 1 to 5% range.
- The total impervious area contributing to any single discharge point shall not exceed 1000 square feet and shall drain to a pervious filter until reaching a property line or drainage swale
- The disconnection shall not cause basement seepage. Normally, this involves extending downspouts at least ten feet from the building if the ground does not slope away from the building

Disconnection with Soil Amendment (50% RR)

- See Soil Restoration Design Specification
- If an amended lawn area does not receive any off-site runoff from impervious surfaces, a 75% runoff reduction can be used.

Disconnection to Rain Garden or Dry Well (50% to 75% RR)

- Depending on soil properties, roof runoff may be filtered in a shallow rain garden or infiltrated into a shallow dry well.
- In general, these areas will require 10 to 15% of the area of the contributing roof area
- An on-site soil test is needed to make the choice of what option to use.
- The facility should be located in an expanded right of way or stormwater easement so that it can be accessed for maintenance.
- For high density sites, front yard bioretention may be an attractive option

Disconnection to Rain Tanks or Cisterns (40% RR)

- The practice for each of these devices depends on their storage capacity and ability to drawdown water in between storms for reuse as potable water, greywater or irigation use.
- Designers will need to estimate the water reuse volume, based on the method of distribution, frequency of use, and seasonally adjusted indoor and/or outdoor water demands for the building
- Based on the prevailing climate for the region, a conservative runoff reduction estimate of 40% is recommended for initial design
- Pretreatment measures may need to be employed keep leaves, bird droppings and other pollutants from entering the tank or cistern
- All devices should have a suitable overflow area to route extreme flows into the

Table D-4. Rooftop Disconnection

next treatment practice or stormwater conveyance system

Table D-5. Grass Channels

Description: The non-roadway portion of the area draining to the grass channel (rooftop, driveway and sidewalk impervious cover and turf cover)

Computation: A 20% reduction in runoff volume is offered for combined turf and impervious cover draining to qualifying swales on A/B soils (10% on C/D soils)

Eligibility: A qualifying grass channel meets the following criteria:

- Primarily serves low to moderate residential development, with a maximum density of no more than 4 dwelling units per acre
- The bottom width of the channel should be between 4 to 8 feet wide. If suitable soil amendments are provided (see Table E-2), the 20% runoff reduction rate may be used for C/D soils
- Swale side-slopes should be no steeper than 3H:1V
- The longitudinal slope of the channel should be no greater than 2%. (Checkdams may be used to break up slopes on steeper swales)
- 5 acres maximum contributing drainage area to any individual grass channel
- The dimensions of the channel should ensure that runoff velocity is non-erosive during the two-year design storm event and safely convey the locals design storm (e.g., ten year design event)
- Designers should demonstrate that the channel will have a maximum flow velocity of less than one foot per second during a one-inch storm event

Note: Where feasible, the dry swale is always the preferable option due to its greater runoff reduction and pollutant reduction capability.

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APPENDIX E		
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